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# CO<sub>2</sub> and H<sub>2</sub>O fluxes due to green manuring under Mediterranean conditions

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**Abstract**. Green manure (GM) is supported by agronomical protocols for improving soil fertility and sustainability of agriculture. The environmental and agronomical efficiency of GM can be evaluated in terms of carbon dioxide (CO<sub>2</sub>) and water (H<sub>2</sub>O) fluxes. These fluxes were monitored by eddy covariance method, together with meteorological data, before and after faba bean (*Vicia faba* spp. minor L.) GM performed in a semi-arid Mediterranean region. The addition of fresh biomass into the soil increases the microbial activity with a positive priming effect: the addition of ready decomposable carbon to the soil increased CO<sub>2</sub> respiration and surface evaporation. The measured C emitted under CO<sub>2</sub> form during the whole experimental period corresponded to 78% of the C incorporated, while evaporated H<sub>2</sub>O in the same period represented 72% of water supplied by encapsulated biomass and precipitations. The cumulated loss of C after GM until the soil preparation of the successive crop was 2957 kgC ha<sup>-1</sup>, in the same period 200 mmH<sub>2</sub>O were lost by soil evaporation.

**Keywords**. Faba bean (*Vicia faba spp. minor L.*), semi-arid, eddy covariance, surface temperature, aridity index.

# 1. INTRODUCTION

The green manure (GM) by using a short-season legume crop is a suitable agro-technique because improves soil organic carbon content (C), sustains productivity and represents an important ecosystem service mainly related to the soil nitrogen (N) recycling (Yadav et al., 2000; Tosti and Guiducci, 2010; Couëdel et al., 2018). The practice is suggested by national and international agronomical protocols, such as the developing plan of rural areas of Italian regions (PSR 2014-2020, Measure 10.1.2) which supports the introduction of faba bean or field peas between two cereal crops, and the European EC Regulation No 834/2007 which imposes GM in organic production (EC, 2007). The ploughing of legume cover crops is usually performed between the flowering and the early pod stages, leaving biomass on the surface or/and incorporating it, whole or chopped, into the soil, ensuring N and C gains for the development of the subsequent crop (i.a. Alluvione et al., 2010). Furthermore, GM increases soil fertility and quality (Zhang and Fang, 2007) and stimulates microbial growth and activity and nutrient mineralization (Eriksen and Gross, 2005). In particular, faba bean (*Vicia faba L.*)

can provide up to 270 kg fixed N ha-1 (Köpke and Nemecek, 2010), including a contribution of 100 kgN ha<sup>-1</sup> from the below-ground biomass (Rochester et al., 1998). However, the application of residues with low C:N such as the legume crops increases the rate of biomass decomposition (De Neve and Hofman, 1996), increasing carbon dioxide (CO<sub>2</sub>) fluxes (Huang, et al., 2004). Muhammad et al. (2019), by a meta-analysis, showed that residues placed at soil surface will reduce CO<sub>2</sub> emissions compared to their incorporated due to a decreasing in the rate of decomposition. The ending of  $CO_2$  losses due to GM is uncertain: the  $CO_2$  emission process can exhausted in a few days (Alluvione et al., 2010; Heller et al. 2010; Shaaban et al., 2016), in one month (Mancinelli et al., 2013) or it can last until several months (Curtin et al., 2000). Surely, when GM implies crop incorporation into the soil, the starting of CO<sub>2</sub> emissions after GM are due to the tillage operation which causes the sudden physical release of CO<sub>2</sub> entrapped in soil pores from previous microbial activity (Reicosky et al., 1997). The duration of this degassing, as defined by Rochette and Angers (Rochette et al., 1999), can be of few hours, depending on soil microclimatic conditions during the time of the year when the tillage is performed (Álvaro-Fuentes et al., 2007; Rochette et al., 1999). Moreover, the  $CO_2$  emissions can be explained by the increased soil organic matter (SOM) decomposition due to the addition to the soil of fresh biomass containing ready decomposable C (Kuzyakov and Domanski, 2000). However, literature reports different outcomes when comparing CO<sub>2</sub> fluxes from bare and GM amended soil, with cases of no differences (Liebig et al., 2010) and others showing-an increasing of CO<sub>2</sub> losses by GM (Kallenbach et al., 2010). These differences can be explained taking into account the plant species and critical high emission moments after cover crop residues addition to the soil (Bodner et al., 2018). Mancinelli et al. (2013) demonstrated that when the green matter is incorporated into the soil, the combined effect of temperature and precipitation, given in terms of aridity index, could cause a priming effect of SOM, also promoting soil mineralization processes and increasing CO<sub>2</sub> fluxes.

Finally, considering the impact of water and temperature on soil CO<sub>2</sub> dynamics, it is necessary to evaluate CO<sub>2</sub> fluxes either before or after GM, as well as to study their relationships with H<sub>2</sub>O fluxes and temperature regime, in order to understand the effects of GM on the environment in any climatic region (Skinner et al., 2019). The continuous monitoring of CO<sub>2</sub> and H<sub>2</sub>O fluxes at hourly scale, together with meteorological variables, immediately before and after GM, gives the chance to identify relationships among CO<sub>2</sub> emissions by GM and the climatic conditions.

The main aim of this study was to verify that GM increased soil  $CO_2$  emissions due to (i) tillage and (ii) input

of organic residues, evaluating the climatic effects on these two sources of  $\rm CO_2$  emissions.

## 2. MATERIALS AND METHODS

## 2.1 The site and the crop

The trial was carried out on a 2.5 ha field located within the CREA-AA experimental farm sited in southern Italy (Rutigliano-Bari, 41° 01' N, 17° 01' E, altitude 147 m a.s.l). Faba bean (*Vicia faba spp. minor L.*) was sowed on 19<sup>th</sup> February 2014 and GM was carried out on 22 May 2014, incorporating crop into the soil at 20 cm depth. Following the usual practice of the area, it could be estimated by eye that almost 70% of the aerial green part of the crop was incorporated into the soil.

The soil was classified as fine, mixed, super-active, thermic Typic Haploxeralfs, clay-loam. Total N and total organic carbon content (TOC) at two depths (0–20 cm and 20–40 cm) were evaluated with mean values equal to 0.92  $\pm$  0.06 gN kg<sup>-1</sup> and 10.3 $\pm$  0.12 gC kg<sup>-1</sup>, respectively.

The climate at the site is classified as semi-arid Mediterranean, with mild winters and warm-dry summers; the mean annual temperature was of 15.7 °C and the mean annual precipitation was of 535 mm year<sup>-1</sup> (Campi et al., 2009), mainly concentrated between the autumn and the late winter. The mean annual water deficit (reference evapotranspiration, ET<sub>0</sub> less rain) was 560 mm.

Fresh and dry weight of above and below-ground biomass, together with plant height and double-sided green leaf area (LAI Licor® 300, USA) were measured according to a randomized sampling design. Mean LAI was equal to  $6.2 \pm 2.2 \text{ m}^2 \text{ m}^{-2}$  and mean height was equal to  $81.8 \pm 16.6$  cm. Above-ground C and N were quantified by TOC Vario Select analyser (Elementar, Germany) and the Kjeldahl method, respectively: the above-ground C and N were 42.2% and 2.7% on a dry weight basis, respectively. The below-ground C and N were equal to 43.2% and 1.6%, respectively. On the basis of the above reported data, the above-ground biomass provided to soil a total estimated supply of 208 kgN ha-1, 3230 kgC ha<sup>-1</sup> and 4.6 mmH<sub>2</sub>O. The below-ground plant parts contributed with an amount of 22 kgN ha<sup>-1</sup>, 590 kgC ha<sup>-1</sup> and 1.5 mmH<sub>2</sub>O, respectively. The C:N ratio was equal to 15.5 and 27.3 for the aboveground and the belowground biomasses, respectively.

# 2.2 $CO_2$ and $H_2O$ fluxes

 $CO_2$  and  $H_2O$  fluxes were measured by eddy covariance (EC) method, a micrometeorological technique

based on the covariance between the instantaneous vertical wind speed (*w*) and the concentration of the scalar (Lee et al., 2004).

The EC CO<sub>2</sub> and H<sub>2</sub>O fluxes were measured in 2014 from March 7<sup>th</sup> until September 17<sup>th</sup>: see Rana et al. (2018) for all details about the trial and eddy covariance data analysis carried out by EddyPro<sup> $\circ$ </sup> software (http://www. licor.com/eddypro). In this study the period from May 15<sup>th</sup> (one week before GM) until September 17<sup>th</sup> (last day before new soil managements for the following crop) was analysed.

The EC tower were equipped with a three-dimensional sonic anemometer (USA-1, Metek GmbH, Germany) for measuring wind components and sonic temperature, and a fast-responding open-path infrared gas analyser LI-7500 (IRGA, LI-7500, Li-COR Inc., Lincoln NE, USA) for detecting atmospheric CO<sub>2</sub> and H<sub>2</sub>O concentrations. These sensors were placed at the centre of the plot at 1.2 m above the top of the canopy, adjusted to follow its growth for the entire growing season, while, after GM, the measurement height was at 1.10 m above soil surface. Data were sampled at 10 Hz by an Industrial Computer (Advantech, USA) running a resident software (MeteoFlux, Servizi Territorio, S.n.c., Cinisello Balsamo, Italy) and recorded at hourly time step.

The Mauder and Foken (2004) and Kormann and Meixner (2001) approaches were followed for performing quality check and footprint analysis, respectively.

#### 2.3 Ancillary measurements and aridity index

Air temperature  $(T_{air})$ , relative humidity, precipitation and incident global radiation  $(R_g)$  were continuously measured at hourly time step by standard sensors placed on a grass meadow close to the experimental field. The soil-crop surface temperature  $T_{surf}$  was measured by four infrared radiometers (IRR-P, Apogee Instruments Inc., Utah, USA) at the centre of the plot, close to the EC tower, positioned 1 m above the crop, at an angle of 60° on the horizontal and pointing toward each of the four cardinal points. Data were collected until 1<sup>st</sup> July by a CR10X (Campbell Scientific Inc., USA) at a frequency of 0.1 Hz and stored for as hourly averages; on 2<sup>nd</sup> July a serious failure interrupted the acquisition of  $T_{surf}$  values.

The aridity index, *AI*, firstly introduced by De Martonne (1926) and widely used to link climatic characteristics and agricultural features (see for example Francaviglia et al., 2017; Emadolin and Reinsch, 2018), was calculated at monthly scale as:

$$AI = \frac{P_i}{T_i + 10}$$

where  $P_i$  = monthly precipitation,  $T_i$  = monthly mean air temperature. Following Mancinelli et al. (2013), the aridity index can be used to assess temporary variations of crop water conditions during crop growing season and can be related to CO<sub>2</sub> flux to verify the relationships between climatic factors and soil respiration.

#### 3. RESULTS AND DISCUSSION

#### 3.1 Meteorological conditions

The weather conditions recorded during the experimental period ( $T_{air}$ ,  $R_g$ , vapour pressure deficit VPD, rain) are reported in Table 1 at monthly scale. In the whole period  $T_{air}$ , VPD and  $R_g$  followed the typical seasonal patterns for this Mediterranean area, with low values of VPD and  $R_g$  recorded during rainy and cloudy days. The precipitation throughout the experimental trial was negligible in the first days after crop incorporation until 14 June, when a huge precipitation event (30.6 mm) occurred; June and July were particularly rainy for this area (see Tab. 1) with precipitations concentrated in few stormy events.

 $T_{surf}$  followed  $T_{air}$  pattern even if about 10 °C above it during the hottest hours of the day and with similar values during the night-(data not shown).

**Tab. 1.** Weather variables at monthly scale: air temperature  $(T_{air})$ , global radiation  $(R_g)$ , vapour pressure deficit (*VPD*), rain: all values are mean, except rain, which is expressed as sum.

Month	$T_{air}$ (°C)	$R_g$ (MJm <sup>-2</sup> d <sup>-1</sup> )	VPD (kPa)	Rain (mm)
May	17.2	20.4	1.4	17.1
June	21.3	24.4	1.8	108.1
July	22.8	21.7	1.8	107.5
August	24.1	21.3	2.0	8.5
September	20.2	15.1	0.8	30.8

**Tab. 2.** Footprint dimensions during daytime and night-time, calculated with the Kormann and Meixner (2001) footprint approach. Length is the length of the footprint, width is the width of the ellipse at the centre of length.

	Daytime		Night time		
	Length (m)	Half width (m)	Length (m)	Half width (m)	
Mean	75.22	9.89	109.99	13.38	
Max	203.84	24.40	547.53	53.85	

## 3.2 Footprint analysis

The analysis of the source area is summarised in Tab. 2. The mean peak emitting point of the flux footprint was estimated to be around 12 m away from the mast, while on average 90% contribution to total fluxes during the daytime was provided by an area extending 134.4 m upwind.

# 3.3 $CO_2$ and $H_2O$ fluxes

On the experimental period (15 May - 17 September) 56, 32, and 12 % of data were flagged 0 (best quality fluxes), 1 (fluxes suitable for general analysis) and 2 (discarded fluxes), respectively (Kormann and Meixner, 2001).

Fig. 1 shows CO<sub>2</sub> and H<sub>2</sub>O fluxes at hourly scale from one week before GM until the end of the monitoring period. Before GM, daily trends of CO<sub>2</sub> fluxes followed the typical pattern of an active green crop with negative daytime photosynthesis values and night-time positive ecosystem respiration values. In the days following the GM, the CO<sub>2</sub> fluxes became positive during daytime, following a typical bell shape pattern; in particular, only soil respiration was detected until the soil remained bare. During the first 6 days immediately after GM the increasing observed in CO<sub>2</sub> emissions was due to the effect of tillage performed for incorporating the faba bean, after that the pattern of CO<sub>2</sub> flux was gently decreasing until the end of the considered period of measurements. During the last three days of monitoring, the CO<sub>2</sub> fluxes showed the pattern of an active green crop since the emerging of weeds started. It is also clear that the CO<sub>2</sub> emissions increased after any effective rain event, mainly due to the restart of decomposition of encapsulated biomass of remaining residue (Muhammad et al., 2019; Curtin et al., 2000), this behaviour was also observed in other studies (Álvaro-Fuentes et al., 2007), and can be explained by the physical release of the  $CO_2$  entrapped in soil pores due to water filling and by the subsequent stimulation of soil microbial activity.

To further investigate the effect of ploughing on  $CO_2$  emissions a zoom on the period immediately before and after GM until the first effective rain (13 June) is reported in Fig. 2, where the components of ecosystem respiration were obtained by the partitioning method applied to the eddy covariance data (Rana et al., 2018). In particular, considering the short-term effect of tillage, few days after GM,  $CO_2$  emission was the result of the above-mentioned *degassing* and the stimulated soil microbial activity by the addition of fresh biomass. The soil respiration increased from 1.76 µmolm<sup>-2</sup>s<sup>-1</sup> (mean over the week before tillage) to 5.46 µmolm<sup>-2</sup>s<sup>-1</sup> (mean over 6 days after GM).

In order to take into account the effect of water on  $CO_2$  fluxes, hourly fluxes of  $H_2O$  were also analysed. Before the GM the water (see Fig. 1) showed the typical daily patterns and absolute values of the actual evapotranspiration of a well-watered crop under Mediterranean climate (Katerji and Rana, 2006). With the termination of the crop by the GM, the  $H_2O$  flux decreased due to plant transpiration suppression and only evaporation process occurring. This clear trend did not occur after the first rain effective event on  $13^{\text{th}}$  June, in fact after this day the  $H_2O$  evaporation from bare soil was addressed by the soil water content following the water supply. Other details, including the overall trends of  $CO_2$  and  $H_2O$  fluxes for this crop along the growing season since plants' emergence, are reported



Fig. 1. Time trends of hourly  $CO_2$  and  $H_2O$  fluxes during the trial. The rainfall and the time interval of harrowing operation for green manure are also indicated.



Fig. 2. Zoom of soil respiration from one week before green manure (GM) until the day of first effective rain event (13<sup>th</sup> June).



Fig. 3. Daily values of  $\rm CO_2$  vs  $\rm H_2O$  fluxes in the first days after green manure until 13th June.

in Rana et al. (2018).

Since both  $CO_2$  and  $H_2O$  fluxes decreased in absolute values starting from the first day after GM until 13<sup>th</sup> June, a possible relationship between these two variables was inves-

tigated. Actually, the comparison between total daily values of  $CO_2$  and  $H_2O$  fluxes (Fig. 3) confirmed a correlation between them due to the analogy in the mechanisms of transport toward the atmosphere of these gases of non-stomatal origin in the absence of rain (Scanlon and Sahu, 2008).

## 3.4 Soil Surface Temperature effect on $H_2O$ and $CO_2$ fluxes

Considering the widely demonstrated relationship between surface temperature and the soil surface evaporation (Kalma et al., 2008; Qiu and Ben-Asher, 2010), the linkage between these variables was investigated at daily scale, always taking into account the period from one week before GM until 13<sup>th</sup> June. Fig. 4 showed a clear decreasing of H<sub>2</sub>O fluxes with the increasing of surface temperature. In particular, two periods can be distinguished: the first one before GM, when the green crop was active, and the second one<del>,</del> after GM, when plants were dead and encapsulated into the soil.

Then, considering that  $CO_2$  emissions were strongly related to soil surface evaporation (see Fig. 3), which depends on surface temperature (see Fig. 4), the relationship between  $CO_2$  emissions and the mean daily surface temperature was further investigated. The daily trend of  $CO_2$ flux until 13<sup>th</sup> June is shown in Fig. 5: it is clear that the  $CO_2$ flux sudden increased after GM for tillage effect. Moreover,  $CO_2$  emissions further increased after rainy days before green manure



Fig. 4. Daily  $H_2O$  fluxes vs surface temperature (*Tsurf*) for the period 15 May - 13 June.

on 26<sup>th</sup> and 30<sup>th</sup> May and 1<sup>st</sup> June. One week before GM the daily CO<sub>2</sub> balance was still negative, and the actively growing crop performed as a CO<sub>2</sub> sink. Just after the GM incorporation, the soil became a net source of CO<sub>2</sub>, reaching a maximum 2 days after the GM (24 May, around 0.5 mol m<sup>-2</sup> d<sup>-1</sup>) and slowly decreasing for several days until it

became almost constant during the last displayed thirteen days. These results are supported by other authors: Mancinelli et al. (2013) reported analogous trends for the peaks of  $CO_2$  emitted immediately after green manuring due to the addition of fresh biomass containing ready decomposable C, which enhances SOM decomposition (Kuzyakov and Domanski, 2000). This positive priming effect results in soil organic matter mineralization and subsequent  $CO_2$  efflux toward atmosphere.

The relationship between CO<sub>2</sub> emissions as daily mean of fluxes ( $\mu$ molm<sup>-2</sup>s<sup>-1</sup>) and  $T_{surf}$  is displayed in Fig. 6 by considering all available  $T_{surf}$  data (from May 23<sup>th</sup> until July 1st). In the first two weeks after incorporation (red symbols) CO<sub>2</sub> emissions were strongly related to  $T_{surf.}$ while during the last monitored weeks (white symbols)  $T_{surf}$  increasing was not related to any increasing in CO<sub>2</sub> emissions. This latter lack of relationship between  $T_{surf}$ and CO<sub>2</sub> fluxes could be explained considering that in this stage CO<sub>2</sub> fluxes likely came from layers of soil below the soil surface where surface temperature did not have effect. This behaviour is similarly to the results reported by Rochette et al. (2000) and Bol et al. (2003) for CO<sub>2</sub> emissions following slurry-C incorporation in grassland. These authors highlighted a two-stage decomposition process: (i) rapid increase in emitted  $CO_2$  due to the decomposition of the labile and easily available fraction of C contained in the biomass, (ii) slow decrease due to decomposition of less available recalcitrant C material. In the present study, the advantage is the chance to follow at hourly scale in continuous the dynamics of these CO<sub>2</sub>



Fig. 5. Daily rain,  $CO_2$  fluxes and surface temperature (*Tsurf*) from one week before green manure (GM) until the first effective rain (13<sup>th</sup> June).



**Fig. 6.** Relationship between  $CO_2$  emissions as means of hourly fluxes at daily scale and mean daily surface temperature.

releases coming from two different sources, highlighting also the effect of tillage.

The strong relationship between surface temperature and  $CO_2$  flux during the first two weeks after GM could suggest that the majority of  $CO_2$  production occurred in plants residue likely before it was completely dry. Parkin and Kasper (2003), in residue covered no-till field crops (corn and soybean) in Iowa (USA), and Rana et al. (2016), in a bioenergetic crop in semi-arid climate in Italy, found similar relations between  $CO_2$  flux and soil surface temperature.

#### $3.5 \text{ CO}_2$ fluxes and the Aridity Index

Fig. 7 displays the cumulated monthly soil CO<sub>2</sub> fluxes (in mol m<sup>-2</sup>) in function of the monthly values of *AI*: the greatest values of CO<sub>2</sub> corresponds to the highest value of *AI*. This result is in agreement with Mancinelli et al. (2013): under Mediterranean conditions, high *AI* value at the time of green manuring may promote the start of soil mineralization processes which may result in a priming effect of native SOM. Thereafter, the decreasing in *AI* corresponds in a decreasing of CO<sub>2</sub> fluxes. In order to develop a finding useful for agronomical purpose, a linear relationship between CO<sub>2</sub> flux and *AI* at monthly scale was investigated (see small upper panel in Fig. 7), founding  $CO_2=0.07 AI + 0.06$  with high value of r<sup>2</sup> (0.68).

## 4. CONCLUSIONS

Eddy covariance measurements of  $CO_2$  and  $H_2O$  fluxes before and after the GM of faba were performed in a typical dry environment of the Mediterranean basin. During the trial, the starting of increased soil metabolic activ-



Fig. 7. Monthly patterns of Aridity Index and  $CO_2$  flux in the experimental period; in the small upper panel the relationship between theses variable is illustrated ( $CO_2=0.07$  AI+0.06 r<sup>2</sup>=0.68).

ity and respiration due to GM was monitored in terms of increasing in  $CO_2$  emissions, highlighting the *degassing* effect due to tillage and the emissions due to decomposition of new fresh biomass following GM: this latter  $CO_2$  emissions was not affected by the surface temperature.

Furthermore, the addition to the soil of fresh biomass containing ready decomposable C increased  $CO_2$  respiration and surface evaporation. The measured C emitted under  $CO_2$  form during the whole experimental period correspond to 78% of the C incorporated, and evaporated  $H_2O$  in the same period represents 72% of water supplied by encapsulated biomass and precipitations.

The cumulated loss of C after the GM until the soil preparation of the successive crop was 2957 kgC ha<sup>-1</sup>; in the same period, 200 mmH<sub>2</sub>O were lost by soil evaporation. This value of emitted C after GM is comparable to those found by Hoyle et al. (2011) which report between 50% and 75% of the C in plant residues can be lost as  $CO_2$  by microbes during the first year of decomposition.

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## REFERENCES

- Alluvione, F. Bertora, C. Zavattaro, L. Grignani, C., 2010. Nitrous Oxide and Carbon Dioxide Emissions Following Green Manure and Compost Fertilization in Corn. Soil Sci. Soc. Am. J., 74(2), 384–395.
- Álvaro-Fuentes, J. Cantero-Martínez, C. López, M.V. Arrúe, J.L., 2007. Soil carbon dioxide fluxes following tillage in semiarid Mediterranean agroecosystems. Soil Till. Res., 96, 331–341
- Bodner, G. Mentler, A. Klik, A. Kaul, A.-P. Zechmeister-Boltenstern, S., 2018. Do cover crops enhance soil greenhouse gas losses during high emission moments under temperate Central Europe conditions? *J. Land Manage. Food Environ.*, 68, pp. 171–187
- Bol, R. Moering, J. Kuzyakov, Y. Amelung, W., 2003. Quantification of priming and CO<sub>2</sub> respiration sources following slurry-C incorporation into two grassland soils with different C content. *Rapid Commun. Mass Spectrom.*, 17, 2585–2590, DOI: 10.1002/ rcm.1184.
- Campi, P. Palumbo, A.D. Mastrorilli, M., 2009. Effects of tree windbreak on microclimate and wheat productivity in a Mediterranean environment. *Eur. J. Agron.*, 30, 220–227.
- Couëdel, A. Alletto, L. Tribouillois, H. Justes, E., 2018. Cover crop crucifer-legume mixtures provide effective nitrate catch crop and nitrogen green manure ecosystem services agriculture. *Ecosyst. Environ.*, 254, 50–59.
- Curtin, D. Wang, H. Selles, F. Zentner, R.P. Biederbeck, V.O. Campbell, C.A., 2000. Legume green manure as partial fallow replacement in semiarid Saskatchewan: effect on carbon fluxes. *Can. J. Soil Sci.*, 80, 499–505.
- De Martonne, E., 1926. L'indice d'aridité Bulletin de l'Association des Géographes Français 9, 3–5.
- De Neve, S. Hofman, G., 1996. Modelling N mineralization of vegetable crop residues during laboratory incubations. *Soil Biol Biochem*, *28*, 1451–1457.
- EC, European Commission, 2007. Council Regulation (EC) No 834/ 2007 of 28 June 2007 on organic production and labelling of organic products and repealing Regulation (EEC) No 2092/91. Official Journal of the European Union L 189: 1–23.

- Emadodin, I., Reinsch, T., 2018. Assessing the impact of land use change on aridification in semiarid land. *Land Degrad Dev*, *29*, 3423–3431.
- Eriksen, J. Gross, L., 2005. Sulphur mineralization-immobilization turnover in soil amended with plant residues. *Soil Biol. Biochem.*, *37*, 2216–2224.
- Francaviglia, R., Di Bene, C., Farina, R., Salvati, L., 2017. Soil organic carbon sequestration and tillage systems in the Mediterranean Basin: a data mining approach. *Nutr Cycl Agroecosyst, 107*, 125–137.
- Heller, H. Bar–Tal, A. Tamir, G. Bloom, P. Venterea, R.T. Chen, D. et al., 2010. Effects of manure and cultivation on carbon dioxide and nitrous oxide emissions from a corn field under Mediterranean conditions. J Environ Qual, 39, 437–48.
- Hoyle, F.C., Baldock, J.A., Murphy, D.V., 2011. Soil organic carbon—role in Australian farming systems. In: Tow P, Cooper I, Partridge I, Birch C (eds) Rainfed farming systems. Springer, Netherlands, pp. 339–361
- Huang, Y. Zou, J. Zheng, X. Wang, Y., Xu, X., 2004. Nitrous oxide emissions as influenced by amendment of plant residues with different C:N ratio. *Soil Biol Biochem*, 36, 973–981. https://doi.org/10.1016/j.soilbio.2004.02.009
- Kallenbach, C.M. Rolston, D.E Horwath, W.R., 2010. Cover cropping affects soil N<sub>2</sub>O and CO<sub>2</sub> emissions differently depending on type of irrigation. *Agriculture, Ecosystems and Environment*, 137, 251–260.
- Kalma, J.D., McVicar, T.R., McCabe, M.F., 2008. Estimating land surface evaporation: a review of methods using remotely sensed surface temperature data. *Surv Geophys*, 29, 421–469.
- Katerji, N. Rana, G., 2006. Modelling evapotranspiration of six irrigated crops under Mediterranean climate conditions. Agricultural and Forest Meteorology, 138, 142–155.
- Köpke, A., Nemecek, T., 2010. Ecological services of faba bean. *Field Crops Res*, *115*, 217–233.
- Kormann, R. Meixner, F.X., 2001. An analytical footprint model for nonneutral stratification. *Boundary-Layer Meteorol.*, 99, 207–224.
- Kuzyakov, Y Domanski, G., 2000. Carbon input by plants into the soil. Review. *Journal of Plant Nutrition and Soil Science*, 163, 421–431.
- Lee, X. Massman, W. Law, B.E., 2004. Handbook of micrometeorology: a guide for surface flux measurement and analysis. In: Atmospheric and Oceanographic Sciences Library. Kluwer Academic Publishers, Dordrecht, The Netherlands, pp. 250.
- Liebig, M.A. Tanaka, D.L. Gross, J.R., 2010. Fallow effects on soil carbon and greenhouse gas flux in Central North Dakota. *Soil Science Society of America Journal*, 74, 358–365.

- Mancinelli, R. Marinari, S. Di Felice, V. Savin, M.C. Campiglia, E., 2013. Soil property, CO<sub>2</sub> emission and aridity index as agroecological indicators to assess the mineralization of cover crop green manure in a Mediterranean environment. *Ecological Indicator*, *34*, 31–40.
- Mauder, M. Foken, T., 2004. Documentation and Instruction Manual of the Eddy Covariance Software Package TK2, vol. 26. Universität Bayreuth, Abt. Mikrometeorologie, Arbeitsergebnisse, 44 pp., ISSN 1614-8916.
- Muhammad, I. Sainju, U. Zhao, F. Khan, A. Ghimire, R. Fu, X. et al., 2019. Regulation of soil CO2 and N2O emissions by cover crops: a meta-analysis. *Soil Till. Res.*, 192, 103–112.
- Parkin, T.B. Kaspar, T.C., 2003. Temperature Controls on Diurnal Carbon Dioxide Flux: Implications for Estimating Soil Carbon Loss. Soil Sci. Soc. Am. J., 67, 1763–1772.
- Qiu, G.Y. Ben-Asher, J., 2010. Experimental Determination of Soil Evaporation Stages with Soil Surface Temperature. Soil Sci. Soc. Am. J., 74, 13–22, doi:10.2136/ sssaj2008.0135
- Rana G. Ferrara R.M. Vitale D. D'Andrea L. Palumbo A.D., 2016. Carbon assimilation and water use efficiency of a perennial bioenergy crop (Cynara cardunculus L.) in Mediterranean environment. Agricultural and Forest Meteorology, 217, 137–150.
- Rana. G. Palatella, L. Scanlon. T.M. Martinelli, N. Ferrara, R.M., 2018. CO<sub>2</sub> and H<sub>2</sub>O flux partitioning in a Mediterranean cropping system. *Agricultural and Forest Meteorology*, 260–261, 118–130.
- Reicosky, D.C. Dugas, W.A. Torbert, H.A., 1997. Tillageinduced soil carbon dioxide loss from different cropping systems. *Soil Till. Res.*, 41, 105–118
- Rochester, I.J. Peoples, M.B. Constable, G.A. Gault, R.R., 1998. Faba beans and other legumes add nitrogen to irritated cotton cropping systems. *Aust. J. Exp. Agric.*, 38 (3), 253–260.
- Rochette, P., Angers, D.A., Côté, D., 2000. Soil Carbon and Nitrogen Dynamics Following Application of Pig Slurry for the 19th Consecutive Year I. Carbon Dioxide Fluxes and Microbial Biomass Carbon. *Soil Sci. Soc. Am. J.*, 64, 1389–1395, doi:10.2136/ sssaj2000.6441389x
- Rochette, P. Angers, D.A., 1999. Soil surface carbon dioxide fluxes induced by spring, summer, and fall moldboard plowing in a sandy loam. *Soil Sci. Soc. Am. J.*, 63, 621–628
- Scanlon, T.M. Sahu, P., 2008. On the correlation structure of water vapor and carbon dioxide in the atmospheric surface layer: a basis for flux partitioning. *Water Resour. Res.*, 44, W10418.

- Shaaban, M. Peng, Q. Hu, R. Lin, S. Zhao, J., 2016. Soil Nitrous oxide and Carbon dioxide emissions following incorporation of above- and below-ground biomass of green bean. *Int. J. Environ. Sci. Technol.*, 13, 179–186. https://doi.org/10.1007/s13762-015-0843-9
- Skinner, C. Gattinger, A. Krauss, M. Krause, H.-M. Mayer, J. van der Heijden, M.G.A. M\u00e4de, P., 2019. The impact of long-term organic farming on soil-derived greenhouse gas emissions. *Sci Rep*, 9, 1702, https:// doi.org/10.1038/s41598-018-38207-w.
- Tosti, G. Guiducci, M., 2010. Durum wheat-faba bean temporary intercropping: Effects on nitrogen supply and wheat quality. *Eur. J. Agron.*, 33, 157–165.
- Yadav, R.L, Dwivedi, B.S Pandey, P.S., 2000. Rice-wheat cropping system: assessment of sustainability under green manuring and chemical fertilizer inputs. *Field Crops Res*, 65, 15–30.
- Zhang, M.K., Fang, L., 2007. Effect of tillage, fertilizer and green manure cropping on soil quality at an abandoned brick making site. *Soil Till Res*, *93*, 87–93.